

# Author Query Form

WILEY

Journal: ECS2

Article: 3560

Dear Author,

During the copyediting of your manuscript the following queries arose.

Please refer to the query reference callout numbers in the page proofs and respond.

Please remember illegible or unclear comments and corrections may delay publication.

Many thanks for your assistance.

**AUTHOR:** Please note that missing content in references have been updated where we have been able to match the missing elements without ambiguity against a standard citation database, to meet the reference style requirements of the journal. It is your responsibility to check and ensure that all listed references are complete and accurate.

Query reference	Query	Remarks
1	<b>AUTHOR:</b> Please confirm that given names (blue) and surnames/family names (vermilion) have been identified correctly.	
2	<b>AUTHOR:</b> Please check the hierarchy of heading levels.	
3	<b>AUTHOR:</b> Mardia and Jupp 2000 has been changed to Mardia and Jupp, 2009 so that this citation matches the Reference List. Please confirm that this is correct.	
4	<b>AUTHOR:</b> As per journal style vectors and matrices should be set in boldface roman. Please check and advise whether we can change vectors 'u, v' to bold.	
5	<b>AUTHOR:</b> Please provide initials for the author 'Michaud and Lesage, personal observation'.	
6	<b>AUTHOR:</b> Reference 'R Development Core Team, 2019.' has not been cited in the text. Please indicate where it should be cited; or delete from the Reference List.	

## Funding Info Query Form

Please confirm that the funding sponsor list below was correctly extracted from your article: that it includes all funders and that the text has been matched to the correct FundRef Registry organization names. If a name was not found in the FundRef registry, it may not be the canonical name form, it may be a program name rather than an organization name, or it may be an organization not yet included in FundRef Registry. If you know of another name form or a parent organization name for a “not found” item on this list below, please share that information.

FundRef name	FundRef Organization Name
DFO Species at Risk program	

# Estimating the proportion of a beluga population using specific areas from connectivity patterns and abundance indices

JEAN-FRANÇOIS OUELLET,<sup>1</sup> ROBERT MICHAUD,<sup>2</sup> MICHEL MOISAN,<sup>2</sup> AND VÉRONIQUE LESAGE<sup>1,†</sup>

<sup>1</sup>Department of Fisheries and Oceans Canada, Maurice Lamontagne Institute, 850 route de la Mer, Mont-Joli, Québec G5H 3Z4 Canada

<sup>2</sup>Group for Research and Education on Marine Mammals, 108, rue de la Cale-Sèche, Tadoussac, Québec G0T 2A0 Canada

**Citation:** Ouellet, J.-F., R. Michaud, M. Moisan, and V. Lesage. 2021. Estimating the proportion of a beluga population using specific areas from connectivity patterns and abundance indices. *Ecosphere* 00(00):e03560. 10.1002/ecs2.3560

**Abstract.** An understanding of habitat use and connectivity is key to preserving wildlife habitat integrity. In this study, we present an innovative approach to habitat connectivity and quantitative assessment of their use that is applicable to any terrestrial or aquatic species. Specifically, we aimed at estimating the proportion of the St. Lawrence Estuary beluga population (Canada) likely to use specific habitats within their summer range from habitat connectivity. To achieve this goal, we used a long-term (~30 yr) observational database of beluga herd movements to derive transit corridor and an index of connectivity among habitats. These were then coupled with relative abundance estimates from aerial surveys for each of these habitats. Specifically, our objectives were to (1) verify the existence and location of transit corridors, (2) examine the effect of environmental factors on movement patterns and corridor routes, (3) describe their use by different age classes, and (4) based on connectivity patterns, determine the proportion of the population likely to transit along these corridors and to be present in each habitat. Our results highlighted the existence of a network of transit corridors among habitats that are connected to a central confluence zone, and expanding with three main branches into the three main regions of their summer range (Upper and Lower Estuary, and Saguenay River). They also highlighted the existence of a spatial segregation of sex and age classes in the population. Previously published aerial survey data indicate that on average, <6% of the population can be found within the limits of each habitat at any given time. Using this information within the framework of habitat connectivity provided a minimum estimate of the proportion of the population likely to visit each habitat, with minimum proportions reaching 45% in habitats of the central confluence zone (Île Rouge area). While we detected no evidence of an influence of bottom depth on corridor routes, we identified a tendency of beluga for moving with surface and ebb tide currents. The application of this approach to beluga and other wildlife populations will contribute to better assessing their potential exposure to natural or human-related stressors.

**Key words:** abundance; animal movement; circular statistics; corridors; focal follows; habitat connectivity; St. Lawrence Estuary beluga.

**Received** 16 September 2020; revised 22 January 2021; accepted 28 January 2021; final version received 6 April 2021. Corresponding Editor: Hunter S. Lenihan.

**Copyright:** © 2021 The Authors. This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

† **E-mail:** veronique.lesage@dfp-mpo.gc.ca

## INTRODUCTION

Preserving habitat integrity is an essential component to strategic conservation planning for wildlife species. High habitat integrity refers to a

state of the environment where animals benefit from the best possible fitness while performing vital activities. In many wildlife species, habitats are composed of concentration areas connected by transit corridors (Roever et al. 2013, Pendoley

1 et al. 2014, Sawyer et al. 2019). Concentration  
2 areas are habitat patches where the animals  
3 spend most of their time and engage in activities  
4 such as foraging, breeding, resting, or socializ-  
5 ing. Transit corridors are used for commuting  
6 between concentration areas on a periodic basis.  
7 Corridors are typically characterized by consist-  
8 ent directional patterns in animal movements.  
9 In a context of high habitat integrity, corridors  
10 are beneficial by providing access to resources  
11 and mates, and allowing migration, dispersal,  
12 and recruitment. Their location may be associ-  
13 ated with a gain in fitness resulting from protec-  
14 tion against predators and disturbance, and an  
15 advantageous energetic cost of utilization (Mon-  
16 teith et al. 2018). In contrast, the location of corri-  
17 dors in low-quality habitats may result in  
18 adverse effects on individual health and  
19 increased risk of injury or mortality if stressors  
20 are located on the trajectory (Henein and Mer-  
21 riam 1990, Hobbs 1992, Sawyer et al. 2019). In  
22 species showing some level of fidelity to a home  
23 range, individuals concentrate their activities  
24 and movements within a subset of their popula-  
25 tion's distribution range. A corollary of this  
26 behavior is that not every individual or popula-  
27 tion segment would be equally exposed to poten-  
28 tial stressors. Understanding habitat connectivity  
29 is thus an essential step for qualifying habitat  
30 integrity.

31 Habitat connectivity is generally investigated  
32 with a spatial matrix representing the landscape  
33 features and habitat patches (Kindlmann and  
34 Burel 2008, Cushman et al. 2013). The corridors  
35 connecting habitat patches across the matrix are  
36 ideally determined from observational data, but  
37 these data are often lacking or insufficient given  
38 that animal movement studies are resource-  
39 intensive. In such cases, the corridors connecting  
40 habitat patches across the matrix are delineated  
41 using a modeling approach where the resistance  
42 of the matrix is tested against the movements of  
43 the target species. Matrix resistance is estimated  
44 according to proxies like the species' habitat pref-  
45 erences or avoidance of landscape features  
46 that hinder movements (Fall et al. 2007, McRae  
47 et al. 2008, Cushman et al. 2013). This framework  
48 offers a wide variety of applications in terrestrial  
49 anthropized landscape settings, and the literature  
50 is flourishing in the field of habitat fragmentation  
51 for terrestrial species. In marine environments, the

biotope is often far less spatially structured, and  
habitat preferences or hindrance associated with  
the landscape features are not as well defined.  
Most marine connectivity studies to date have  
dealt with larval dispersion, and comparatively,  
little effort has been devoted to delineating corri-  
dors or assessing habitat connectivity in marine  
megafauna (Bryan-Brown et al. 2017).

The beluga (*Delphinapterus leucas*) population  
from the St. Lawrence Estuary (SLE), Canada, has  
been depleted by commercial and bounty hunting,  
and is currently small (<900 individuals), endan-  
gered, and declining (Mosnier et al. 2015). It is  
exposed to multiple anthropogenic stressors, nota-  
bly contamination from toxic substances, noise pol-  
lution from vessel traffic and marine infrastructure  
development, and small vessel collision risk  
(COSEWIC 2014). The population's summer range  
is downstream of a heavily industrialized water-  
shed and overlaps a busy seaway. Areas within  
their summer range where a large proportion of  
the population gather (50% kernel densities) or  
where beluga movements are less directional have  
been documented by multiple investigators using  
different datasets and are highly consistent among  
studies (Michaud 1993, Lefebvre et al. 2012, Mos-  
nier et al. 2016, Fig. 1). Current knowledge indi-  
cates ample movement among these areas. For  
instance, residency time of individual beluga  
within these areas is notably short, lasting a few  
hours only, with considerable time spent moving  
between them (Chadenet 1997, Lefebvre et al.  
2012). This information suggests a high turnover of  
individual beluga within each area. The beluga is  
also well known for its spatial segregation between  
sexes and age classes at least during summer  
(Smith et al. 1994, Loseto et al. 2006), a characteris-  
tic also documented in the SLE (Michaud 1993).  
Genetic and behavioral evidences indicate high site  
fidelity to general summering areas in beluga, as  
well as to specific habitats within them (Caron and  
Smith 1990, Turgeon et al. 2012, Colbeck et al.  
2013, O'Corry-Crowe et al. 2018, 2020). A better  
understanding of connectivity patterns among SLE  
beluga of different age and sex classes has impor-  
tant implications as it can reveal the proportion of  
a population occurring in specific areas and that is  
potentially exposed to stressors impacting any one  
of these areas or connecting corridors.

Using a long-term database of SLE beluga herd  
summer movements coupled with spatially

Fig 1



1 explicit aerial survey results on population distri-  
2 bution among concentration areas, we aimed at  
3 investigating how these areas are interconnected.  
4 More precisely, our objectives were (1) to verify  
5 the existence and location of transit corridors, (2)  
6 to describe their use by different age- and sex  
7 classes, (3) estimate the proportion of the popula-  
8 tion likely to use each area from connectivity pat-  
9 terns, and (4) test the effect of environmental  
10 factors on herd movements and corridor routes.  
11 The originality of our study stems in the combi-  
12 nation we made of observational data of herd  
13 movements and aerial survey results on distribu-  
14 tional data.

## 15 METHODS

### 16 *Study area and data collection*

17  
18 Connectivity studies require information on  
19 three spatial components: a landscape matrix,  
20 concentration areas (high-use areas or habitat  
21 patches), and corridors. The landscape matrix  
22 used in this study was a grid with 1-km<sup>2</sup> cells.  
23 This cell dimension is consistent with Lefebvre  
24 et al. (2012) who analyzed a subset of the same  
25 data for a different purpose. Concentration areas  
26 (CA; hereafter referred to as such) corresponded  
27 to the 36 areas published in Mosnier et al. (2016)  
28 for the SLE beluga population (Fig. 1). These  
29 CAs were based on 35 systematic line transect  
30 aerial surveys of their entire summer range. They  
31 were delineated with the kernel method and rep-  
32 resent areas where 50% of the observations were  
33 cumulated during these surveys ( $K_{50}$ ). They are  
34 each conveniently associated with a proportion  
35 of the population (summing to 50% not 100%  
36 given they are  $K_{50}$ ), which was assumed to be  
37 1100 individuals at the time of their study  
38 (Appendix S1: Table S1). The Saguenay River  
39 represents an important component of the SLE  
40 beluga summer range, but was not included in  
41 Mosnier's study given that counts were obtained  
42 using a different method, that is, they were  
43 assumed to be total counts (Gosselin et al. 2017).  
44 The proportion of the population using the  
45 Saguenay River was obtained using visual  
46 ( $n = 36$ ) and photographic ( $n = 8$ ) counts for the  
47 Saguenay River extracted from Gosselin et al.  
48 (2017) and assuming a population size of 1100  
49 individuals as in (Mosnier et al. 2016). On aver-  
50 age, an estimated 1.9% and 1.6% (mean 1.8%) of  
51

the population was present in the Saguenay  
River according to visual and photographic sur-  
veys, respectively.

The existence of transit corridors and their  
routes were examined using observational data  
of herd movements obtained from vessel-borne  
focal follows of beluga herds. Data collection  
spanned 28 yr (1989–2016) and was conducted  
on average 2.4 times a week between June and  
October, inclusively, in most years. Most of the  
data were collected by the Group for Research  
and Education on Marine Mammals (GREMM)  
and a second team (DFO, Mont-Joli, Quebec,  
Canada) contributed to sampling in 2003, 2005,  
2010, and 2011. Details of data collection are  
described in Lefebvre et al. (2012). Briefly, selec-  
tion of daily survey area was not random nor  
systematic, but rather made according to  
weather conditions prevailing in different parts  
of the summering area. Caution was taken to  
maximize spatial coverage within the summer  
distribution range and to prevent resampling the  
same area on consecutive days. However, most  
of the search effort was allocated to the central  
portion of the SLE beluga summer distribution  
(Mosnier et al. 2010, Gosselin et al. 2014, 2017),  
within a 35 km radius from the Saguenay River  
Mouth and main port in Tadoussac, and  
decreased from there on with increasing distance  
(Fig. 2). In the Saguenay River, effort remained  
high up to Baie Sainte-Marguerite, but was low  
past this point where beluga occurrence is much  
lower (Gosselin et al. 2014, 2017). All segments of  
the population occur on a regular basis in the  
sampled portion of the summer range (Michaud  
1993); thus, no systematic bias was anticipated  
from the sampling scheme. Search effort was not  
systematically recorded in the early years of this  
program and thus could not be used here to dis-  
tinguish areas with an absence of effort from  
those with effort but absence of beluga (see also  
Lefebvre et al. 2012).

A herd was defined as an assemblage of  
groups of individuals where inter-group distance  
was small relative to herd extent. Groups were  
defined as assemblages of beluga less than one  
body length apart from one another. During focal  
follows, point sampling sessions recorded herd  
location and surface configuration including  
herd size, radius, and age composition, and were  
made at 30-min intervals (see Lefebvre et al.

Fig 2

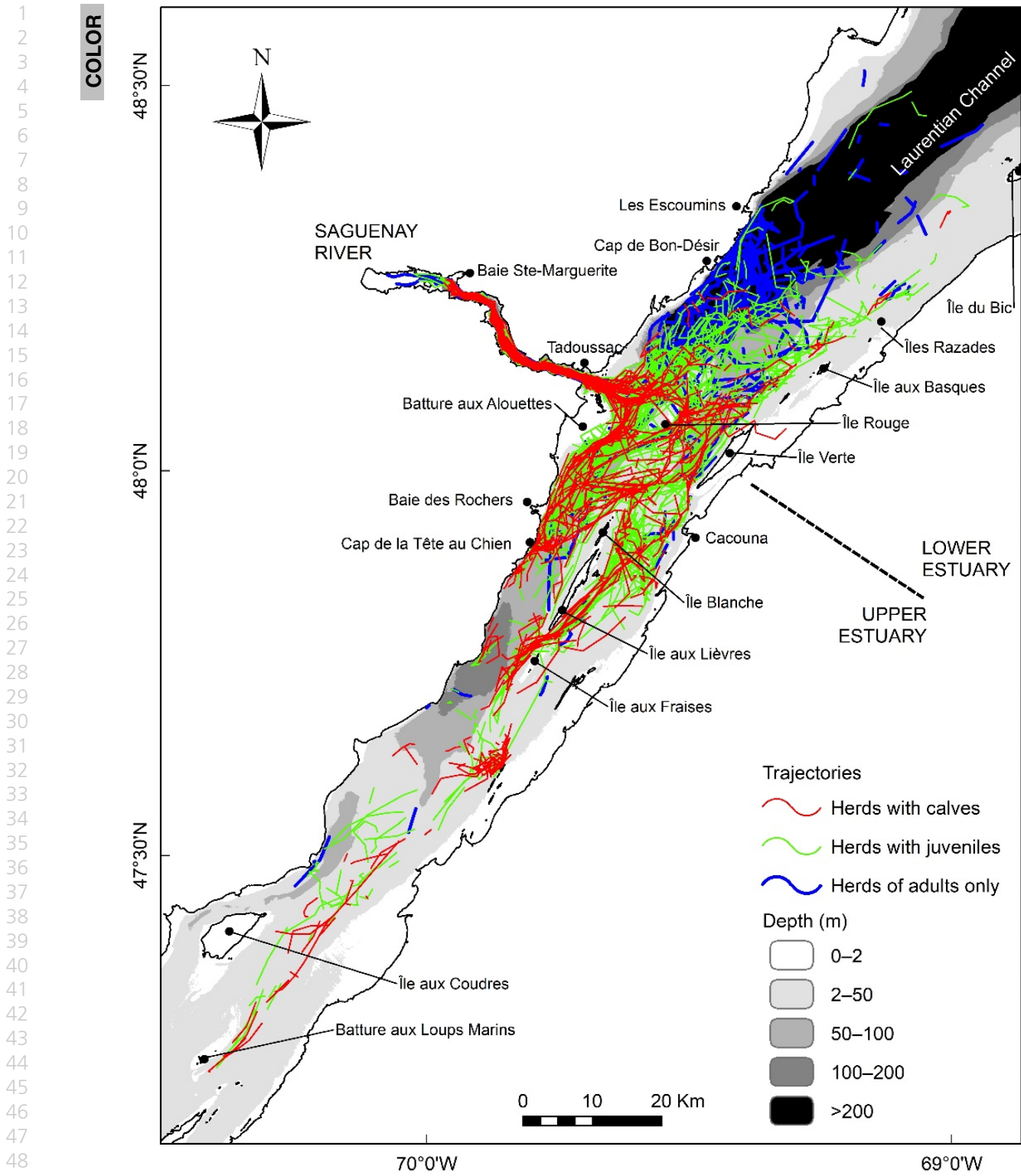


Fig. 2. Beluga herd trajectories from vessel-borne visual follows with a segregation by age classes. Depth data are from Canadian Hydrographic Service.

2018, for the full set of sampled herd characteristics). Upon detection of a herd, a 15-min period of observation from outside the herd was undertaken, after which the vessel approached at slow speed, took position within the herd, and observers proceeded with the first point sampling session. Keeping pace with the animals, the vessel remained within the herd for an average of 93 min (SD = 59; range 548) or until the contact was lost. The location recorded during the point sampling sessions was that of the vessel. Age composition consisted in the estimation of numbers of individuals in age classes with the following bins: calf (small-sized, beige, or dark-colored individuals, jerky surfacing and fetal fold if apparent), juveniles (medium-sized, gray-colored individuals), and adults (white-colored individuals). Male and female beluga cannot be reliably differentiated in the field, and thus, sex composition was not recorded.

#### Data analysis

Given the large amount of observational movement data available for SLE beluga, and the challenges we would have faced estimating resistance matrix in this spatially unstructured biotope where cost-associated habitats (other than very shallow waters and islands) are undefined, application of methods typically used in data-poor connectivity assessments was unnecessary. Instead, we proceeded directly with the delineation of the corridors by a spatially explicit statistical analysis of aggregated herd movement data. Spatial data treatment and analysis was conducted in ArcGIS (ver. 10.4; ESRI, Redlands, California, USA). Only locations obtained from a GPS (2005 and onward) or Loran-C receiver (1989–2007) or those recovered from marine charts according to accurate field notes (1989 and onward) were considered valid. Herd locations recorded in decimal degrees were converted into UTM coordinate system to facilitate subsequent computations. The 30-min interval rule between point sampling sessions was not always strictly followed in the field. Any gap between two point sampling sessions longer than 60 min terminated a follow. Throughout this paper, step refers to an interval, in time and space, between two point sampling sessions. The data collected during a point sampling session were associated with the step it terminated. Step length corresponds to the

Euclidean distance between two point sampling sessions and was calculated according to the Pythagorean theorem. Step direction of movement was calculated relative to true north with trigonometric functions. When two successive point sampling sessions were on either side of a headland, the resulting step crossed over land. In those cases, the step was manually forced to detour the headland with the shortest route by adding a location splitting the step in two parts. The time associated with this artificial location was calculated based on its relative distance between the adjoining locations, and assuming a constant travel speed. Most such occurrences were found in the Saguenay River. Spatial buffers were generated around the steps (using the herd radius recorded in the field), to account for herd spread. Data associated with each step (direction of movement and configuration) were assigned to the cells intersecting their respective buffers.

The verification of the existence of corridors and their delineation were achieved with two methods: vectors of movement and probability of orientation. Vectors of movement were produced in each matrix cell with mean step direction as an index of vector direction and resulting vector length ( $\rho$ ) as an index of vector amplitude.  $\rho$  is a measure of dispersion for circular data. This parameter conveniently scales from 0 to 1 where the latter indicates absence of dispersion (Mardia and Jupp 2009). The distribution of the direction data in each cell was compared to a uniform distribution with Rao's test for homogeneity, Kuiper's one-sample test of uniformity, and Watson's goodness of fit test (circular package in R). Only cells with a minimum number of 10 steps, that is, with a minimum sample size of 10 beluga herds, were tested ( $n = 1153$ ). A uniform distribution would indicate that, in a particular cell, herds were likely to move in any direction. In contrast, a heterogeneous distribution would indicate a directional pattern. A visual examination of the circular diagrams produced for those cells revealed anything between a unimodal distribution to a sharp multimodal distribution (herds likely to move in more than one direction, sometimes in strict opposite direction). In the latter cases, direction data could obviously not be aggregated in a single mean value. Therefore, the cells with heterogeneous

1 distribution were included in a modal analysis  
2 for circular data (circular package in R). Because  
3 of the circular nature of direction data, there was  
4 a risk that a mode that overlapped the point  $0^\circ$   
5 was split in two, artificially inflating the rho  
6 parameter. One way to prevent this possibility  
7 was to rotate the data so that the point  $0^\circ$  fell out-  
8 side the modes. The direction data were therefore  
9 rotated eight times by one-eighth of a circle ( $0^\circ$ ,  
10  $45^\circ$ ,  $90^\circ$ ,  $135^\circ$ ,  $180^\circ$ ,  $225^\circ$ ,  $270^\circ$ , and  $315^\circ$ ) each  
11 time. After each rotation, a modal analysis for  
12 circular data was conducted, the data points  
13 were divided into the resulting modes, and rho  
14 was calculated for each mode. For each cell, the  
15 modes kept for further analysis were the ones  
16 produced by the rotation that issued the smallest  
17 number of modes and the highest cumulative  
18 rho value. In cells with heterogeneous data distri-  
19 bution, vector direction and amplitude (rho)  
20 were calculated for each mode in each cell with  
21 the un-rotated data. In cells where data distribu-  
22 tion was uniform or not tested, vector direction  
23 and amplitude were calculated using data aggre-  
24 gated in a single vector. In order to compare our  
25 results with what can be expected by chance  
26 alone, this procedure was repeated with simu-  
27 lated data where random direction values were  
28 generated in each matrix cell in equal number to  
29 its number of observed steps. In order to allow  
30 contiguous vectors to graphically form parallel  
31 patterns, vector directions were subsequently  
32 assigned an octant number according to its direc-  
33 tion of movement rounded to the nearest multi-  
34 ple of  $45^\circ$ . In this case, each octant pointed  
35 toward the centroid of one of the eight neighbor-  
36 ing cells, with octant 1 being centered at  $90^\circ$ , and  
37 numbers proceeding clockwise from octant 1 to  
38 octant 8, which was centered at  $45^\circ$ .

39 In order to ensure some robustness in corridor  
40 delineation, we used a second approach where  
41 we conducted a hierarchical cluster analysis using  
42 Euclidean distance and complete linkage as the  
43 agglomeration method, where the matrix cells  
44 were grouped according to their probability of  
45 movement along each one of four axes. This prob-  
46 ability was calculated with the following proce-  
47 dure: Step directions were rounded to the nearest  
48 multiple of  $45^\circ$  and assigned the corresponding  
49 octant value. Octants were binned in four axes of  
50 movement as follows: N-S, NE-SW, E-W, SE-NW.  
51 An axis of movement differs from a true direction

in that it assumes complete bi-directionality. For  
example, two steps going at  $45^\circ$  and  $180^\circ$  were  
assigned to the NE-SW and N-S axes, respec-  
tively. Then for each cell, we calculated the pro-  
portion of steps moving along each axis relative  
to the total number of steps in the cell. Only  
matrix cells with 5 steps or more were included.  
The optimal number of clusters was assessed  
with the following methods and using packages  
factoextra and NbClust in R: average silhouette  
width (Kaufman and Rousseeuw 1990), gap  
statistics (Tibshirani et al. 2001), Calinski and  
Harabasz index (Calinski and Harabasz 1974),  
and total within sum of square.

Patterns in connectivity using the vector  
method were examined from maps where the  
degree of directionality for each cell (rho) was  
assessed using a varying level of severity. Vectors  
were considered strong in a cell when direction  
data distribution was heterogeneous and based  
on at least five steps, when amplitude (rho) was  
0.70 or higher, and when the cell was contiguous  
to at least one neighboring vector with the same  
characteristics. Vectors were considered weak  
otherwise. The amplitude condition of 0.70 corre-  
sponds to a range equal to  $45^\circ$  (which is the  
width of an octant and our rounding uncer-  
tainty) with a sample size of 2. Under a high level  
of severity, a corridor had to be based on strong  
vectors showing a parallel pattern. A low level of  
severity allowed gaps in contiguity and vector  
amplitude lower than 0.70. The connective value  
of each corridor was determined by summing  
the averaged proportion of the population asso-  
ciated with every CA it connected and where  
each CA was assigned to only one corridor.

The existence of a match between herd move-  
ment direction and surface current direction was  
examined by assigning to each point sampling  
session the tidal phase and surface current condi-  
tions (direction and speed) that prevailed at the  
time of the observation. Tidal phases were  
extracted from water level data referenced at  
Tadoussac and provided by the *Canadian Hydro-  
graphic Service* (<http://www.tides.gc.ca/eng/find/zone/2>). Data were aggregated in 3-h bins (low,  
rising, high and ebb tides) where high tide occurs  
from 5 to 7 h after low tide. Surface current data  
were obtained from a model (Saucier et al. 1999,  
Saucier and Chassé 2000) predicting planar vec-  
tors (u, v) of surface current speed with a time

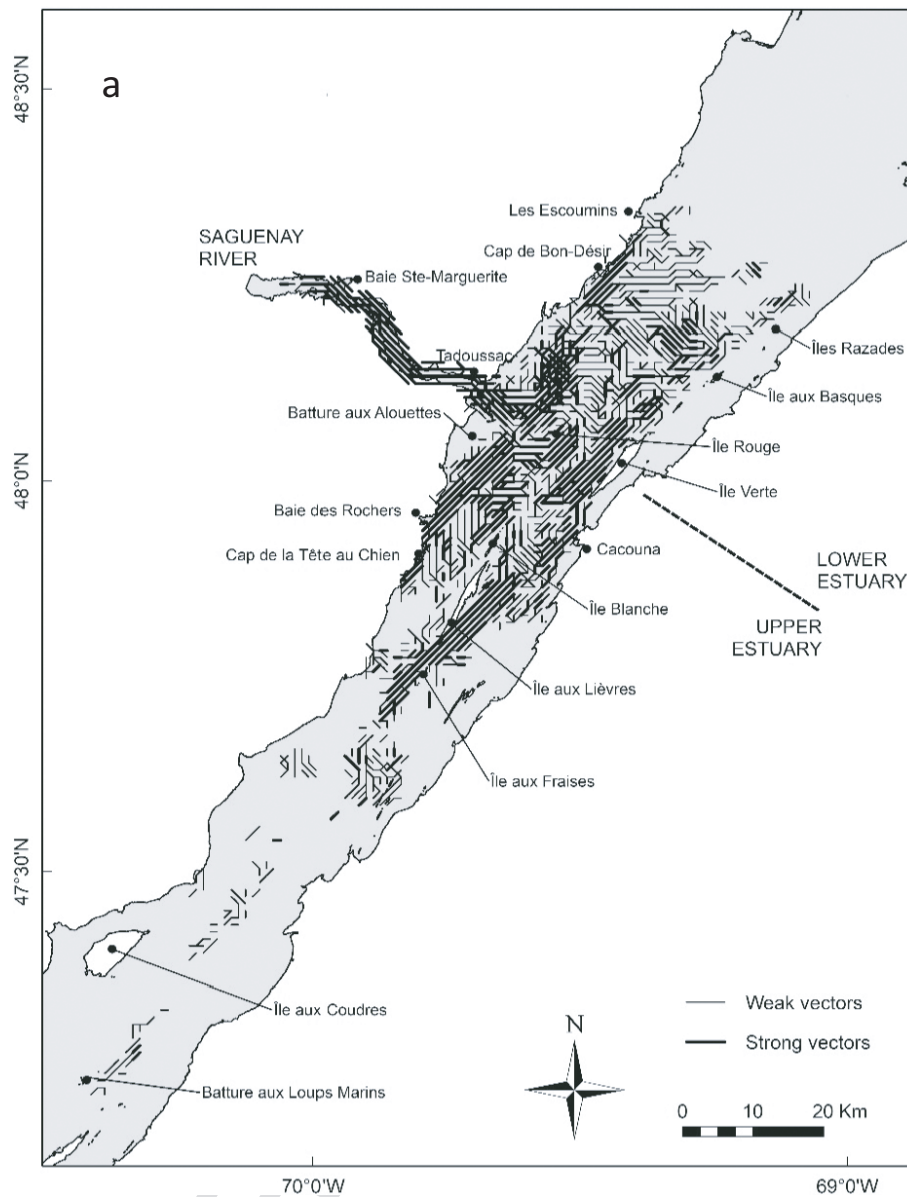


Fig. 3. Direction and amplitude of vectors of movements based on (a) beluga herd trajectories and (b) random directions from simulated movement data as if movements were driven by chance alone.

and spatial resolution of one hour and  $400 \times 400$  m pixels, respectively. A spatial buffer was generated around each point sampling session using herd radius recorded in the field. Each point sampling session was assigned the mean planar vector values averaged over every pixel contained in its buffer. Resulting current speed was calculated with those mean planar vectors according to the Pythagorean theorem. Resulting

current direction was calculated relative to grid north using trigonometric functions.

The influence of current direction on directionality of movements was assessed at two time and spatial scales. The concordance between step direction and current direction was first verified globally using any steps associated with strong and contiguous vectors ( $\rho > 0.70$ ). This was done separately for rising and ebb tides (when currents

(Fig. 3. *Continued*)

are expected to be the strongest), and separately for the SLE and Saguenay River. To account for potential local differences in the influence of currents on movement directionality, the absolute difference between step direction and current direction was calculated (range =  $0^{\circ}$ – $180^{\circ}$ ) and averaged over each strong vector. Data from all tidal phases were examined, but only strong and contiguous vectors were included.

The water depth data (resolution 20 m) were obtained directly from the Canadian Hydrographic Service. Each cell centroid was assigned its corresponding depth value through a spatial join. Water depth was compared with a *t*-test between cells belonging to a corridor and cells not belonging to a corridor. Cells belonging to a corridor were those that contained at least one strong vector, in accordance with our highest

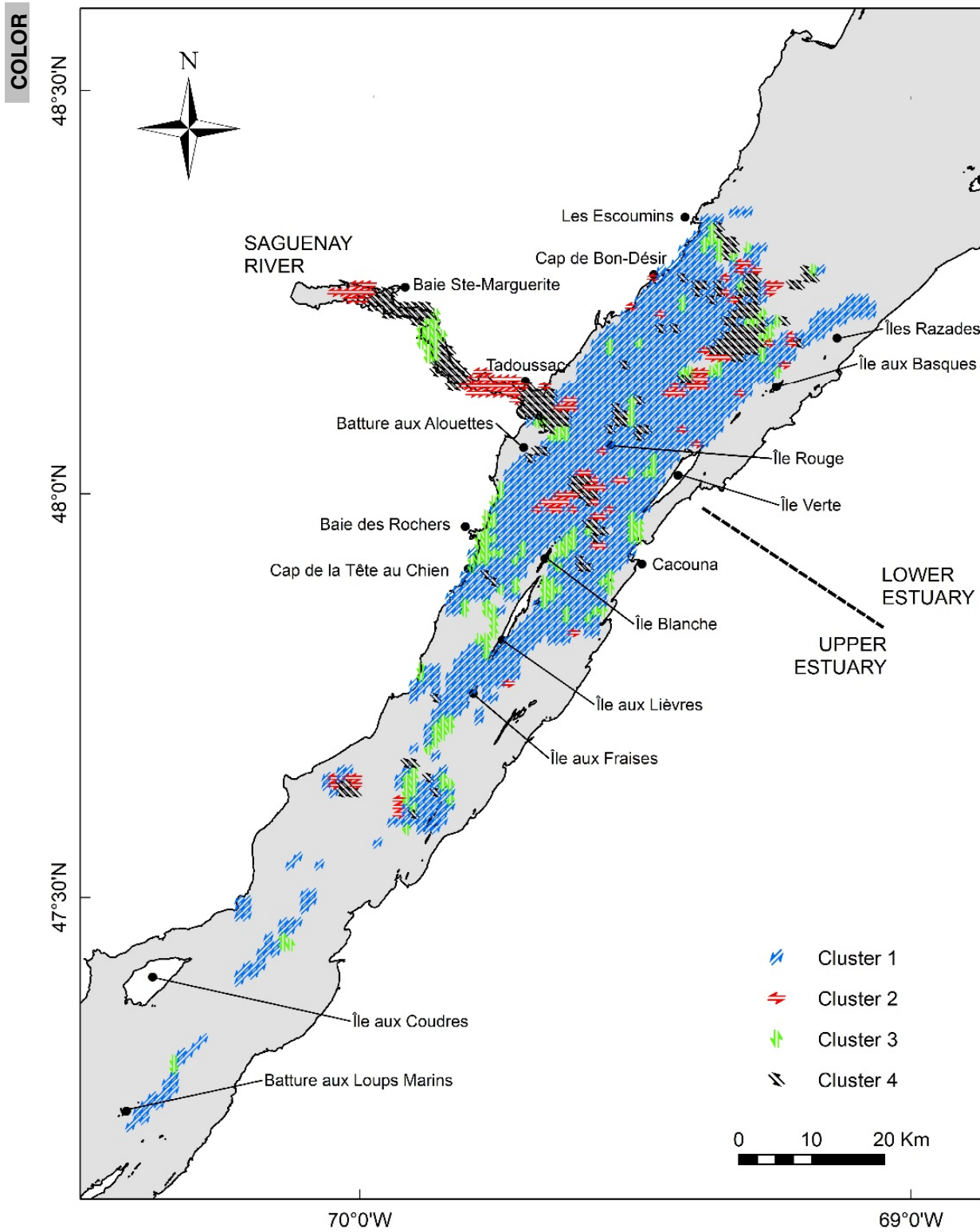


Fig. 4. Spatial distribution of clusters based on probability of orientation of herd movements.

1 level of severity and ignoring the parallelism  
2 aspect. This analysis included only the cells con-  
3 taining a minimum of 5 steps, to allow the possi-  
4 bility of a strong vector. Water depths <2 m were  
5 excluded due to possible navigational constraints  
6 for beluga. The Saguenay River was excluded  
7 from this analysis because its narrowness  
8 (<2 km) caused it to be almost completely cov-  
9 ered with strong vectors, leaving no possibility  
10 for the comparison.

## 11 RESULTS

12  
13  
14 Of the 2897 herd follows conducted between  
15 1989 and 2016, 2224 provided two or more point  
16 sampling sessions with valid location data, for a  
17 total of 9127 point sampling sessions, and an  
18 average of 4 valid point sampling sessions per  
19 follow. The spatial distribution of herd follows  
20 (Fig. 2) reflects in general terms the distribution  
21 of search effort, with a concentration within  
22 approximately 35 km from the Saguenay River  
23 mouth (see *Methods*). Blank areas in the extreme  
24 ends of the study area, in the Estuary and the  
25 Saguenay River, need to be interpreted with cau-  
26 tion as they may not reflect a true absence of bel-  
27 uga but a lack of effort.

28 The steps that compose herd trajectories are  
29 Euclidean lines joining pairs of locations, and the  
30 locations are those of the vessel moving within  
31 the herd. As a result, successive steps of a herd  
32 trajectory may comprise some artificial variabil-  
33 ity. Nevertheless, the distribution of herd trajec-  
34 tories highlights some notable spatial patterns  
35 (Fig. 2). First, the trajectories showed a fluctuat-  
36 ing level of parallelism across the study area.  
37 Notable examples of a sharp parallel pattern  
38 among trajectories were observed in the Sague-  
39 nay River, along the south shore of Île aux  
40 Lièvres, north shore of Île Verte and Batture aux  
41 Alouettes. Examples of a contrasting lack of par-  
42 allelism among trajectories can be found off  
43 Cacouna and near the Saguenay River mouth.  
44 Second, a gyre-like pattern taking the shape of a  
45 roundabout, with many trajectories branching  
46 off to multiple directions, circled around Île  
47 Rouge bank. This feature appeared as a major  
48 confluence zone in the summer range of the SLE  
49 beluga population. Finally, herd compositional  
50 data associated with each trajectory highlighted  
51 an age-specific spatial distribution in beluga

movements in the Estuary, but not in the Sague-  
nay River (Fig. 2).

Combining steps into vectors of movement  
resulted in a total of 2016 vectors comprised of 5  
steps or more, half of which (947) were consid-  
ered strong vectors (Fig. 3a). Examining paral-  
lelism among vectors reinforced patterns  
observed at the level of individual trajectories  
and steps in Fig. 2. Parallel patterns among  
strong vectors could be found in multiple places  
in the study area, notably in the four areas indi-  
cated above, that is, in the Saguenay River, along  
the south shore of Île aux Lièvres, the north shore  
of Île Verte and along the Batture aux Alouettes.  
In most instances, these sets of strong vectors  
were oriented parallel to the coastlines. The ran-  
domly generated data showed multiple strong  
vectors but no parallel pattern (Fig. 3b), indicat-  
ing that chance alone was unlikely to explain pat-  
terns observed in beluga movement orientations.

The average silhouette width method, the gap  
statistic method, and the Calinski and Hara-  
basz's index applied to the probability of orien-  
tation in different directions indicated optimal  
numbers of clusters of 2, 4, and 6, respectively.  
The total within sum of square method was  
inconclusive. We chose a number of four clusters  
as an average point. This cluster-based approach  
showed less spatial structure in beluga move-  
ment orientation than the one using vectors and  
was therefore less informative (Fig. 4). Never-  
theless, the most notable result was the domi-  
nance of cluster 1 with 70% of the cells  
( $n = 1077$ ). This cluster grouped cells charac-  
terized by a northeast-southwest axis of move-  
ment, that is, within the main flowing axis of  
the Estuary, and was distributed exclusively in  
the Estuary (i.e., none in the Saguenay River).  
The other three clusters were found in the Estu-  
ary and the Saguenay River. Cluster 2 was the  
smallest one with 117 cells. It was dominated by  
an east-west axis of movement. Aggregations of  
cells belonging to cluster 2 were found in Sague-  
nay River and near the eastern end of Île  
Blanche in the Estuary. Cluster 3 and 4 showed  
a nearly similar number of cells (173 and 191)  
and were characterized by a north-south and a  
northwest-southeast axes of movement, respec-  
tively. Notable aggregations of cells from cluster  
3 were found in Saguenay River, along Île  
Blanche, along a stretch of coastline between



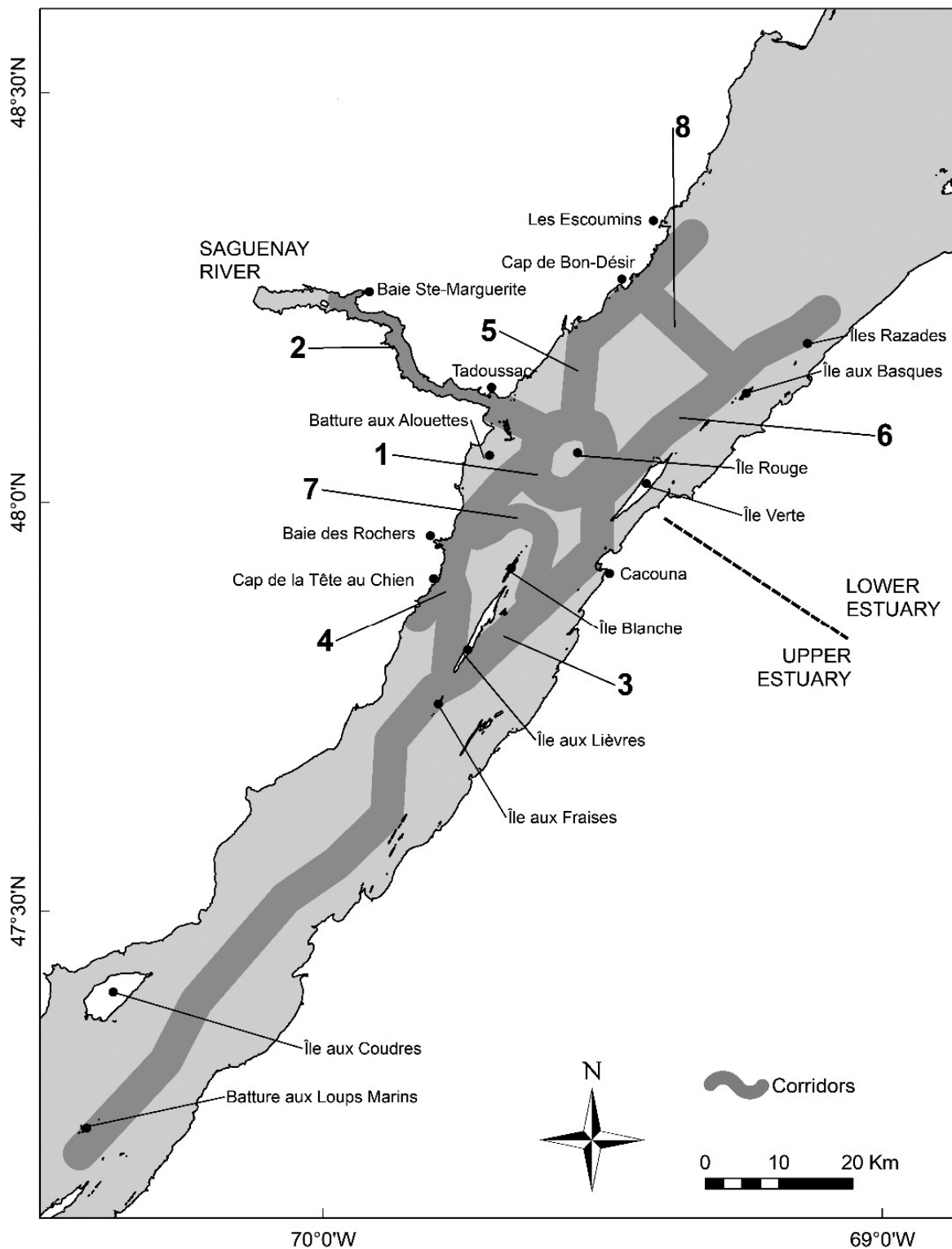


Fig. 5. Network of corridor routes based on observational data of beluga herd movements. Gray lines indicate the shape and location of the corridors, but the actual width and endpoints of the corridors are unknown.

Table 1. Connective value for each transit corridor delineated in the study area.

Corridor	Connected CAs	Connective value	Connected corridors
1	20, 24	5.6	2, 3, 4, 5, 6
2	23, Saguenay River	3.0	1
3	12, 13, 18, 19, 22 (1, 2, 3, 4, 5, 6, 7, 9, 11)	12.1 (18.3)	1, 7 (4)
4	15, 17, 21 (14)	5.7 (6.3)	1, 7 (3)
5	25, 27, 29	3.4	1, 8
6	26, 28, 30 (31, 32)	7.5 (8.5)	1, 8
7			3, 4
8			5, 6

Notes: Values represent the sum of the percentage of the population associated with each concentration area (CA) connected by each corridor. Parentheses show the connective value and connected CAs and corridors when accepting a lower level of severity in corridor delineation. Corridors 7 and 8 do not directly connect any CA. Summary of corridors location: (1) Île Rouge bank, (2) Saguenay River, (3) Upper Estuary, south shore, (4) Upper Estuary, north shore, (5) Lower Estuary, north shore, (6) Lower Estuary, south shore.

Batture aux Alouettes and Cap de la Tête au Chien. Aggregations of cells belonging to cluster 4 were found in Saguenay River, between Île aux Basques and Cap de Bon-Désir and between Île Verte and Batture aux Alouettes, reflecting movements across the Estuary.

The level of connectivity among the CAs was high with eight corridors emerging from the spatial patterns of vectors or cluster aggregations (Fig. 5). A detailed description of these corridors, including vector strength, clusters that support them, CAs that these corridors connect, and type of herds using them is presented in Appendix S2. In summary, corridor 1 circled around Île Rouge and appeared to be a central confluence zone. It was connected to five other corridors (2, 3, 4, 5, 6) extending into the Upper and Lower Estuary and into the Saguenay River. These five corridors were roughly parallel to the coasts. Two other corridors (7, 8) were stretched across the Upper and Lower Estuary (Fig. 5). Generally, corridors extending in the Upper Estuary were seldom used by herds of adults only, while those extending over the Laurentian Channel in the Lower Estuary were mainly used by herds of adults only, and to a lesser extent herds including juveniles (Fig. 2). Based on survey data about the relative distribution of beluga, the results on connectivity among concentration areas indicate that a CA could be

visited by anywhere between 3.0 and 12.1% of the population, depending on CAs when considering each of the six delineated corridors separately (Table 1). The maximum value rose to 18.3% when accepting a lower level of severity. The highest connective values were associated with corridor 3 and 6 running near the southern shore in the Upper and Lower Estuary, respectively. The lowest connective value was in corridor 2 in the Saguenay River. Corridors 7 and 8, running across the Estuary, had no associated connective value because there were no CA on their route. CAs near Île Blanche and Île aux Fraises are those likely to be used by the largest proportions of the population. Due to its central place and to the multiple corridors it connected, the connective value of corridor 1 around Ile Rouge can be increased to 37.3% if we add the connective values of each corridor it connected. Accepting a lower level of severity, this value is further increased to 45.1% and these estimates are probably minimum values since they were based on 50% kernel densities.

Step direction data belonging to strong vectors and recorded during rising and ebb tides showed a bimodal distribution in both the Estuary and the Saguenay River (Rao', Kuiper's, Watson's tests,  $P < 0.01$ ; Fig. 6). Surface current direction data were multimodal at rising tide and unimodal at ebb tide (Rao', Kuiper's, Watson's tests,  $P < 0.01$ ). In the Estuary, mean modal step directions matched the direction of surface current and the orientation of the Estuary in the study area, that is,  $226 \pm 62^\circ$  ( $n = 819$ ) and  $48 \pm 24^\circ$  ( $n = 195$ ) at rising tide, and  $65 \pm 52^\circ$  ( $n = 634$ ) and  $245 \pm 38^\circ$  ( $n = 272$ ) at ebb tide (Fig. 6). In the Saguenay River, mean modal step directions generally also followed the direction of surface current and the orientation of the River (see Fig. 2), and were  $305 \pm 30^\circ$  ( $n = 318$ ) and  $129 \pm 30^\circ$  ( $n = 164$ ) at rising tide, and  $311 \pm 33^\circ$  ( $n = 166$ ) and  $128 \pm 34^\circ$  ( $n = 151$ ) at ebb tide.

Our analysis examining the data at fine scale to identify local peculiarities of movement-current associations revealed ample consistency between herd and surface current directions (Fig. 7). The average difference between movement and current directions for the 968 strong vectors included in the analysis was mostly comprised between  $45^\circ$  and  $90^\circ$  and, to a lesser extent, within less than  $45^\circ$  of each other.

Table 1

Fig 6

Fig 7

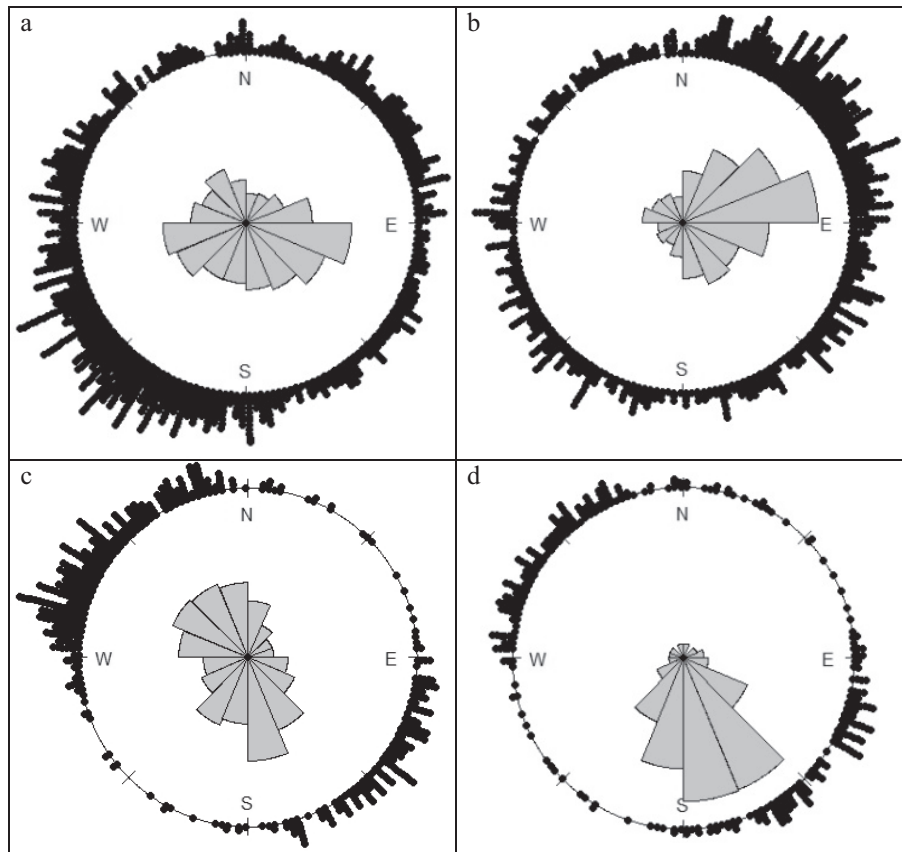


Fig. 6. Step directions associated with strong and contiguous vectors (dots) and surface current directions simultaneous with either one of the steps (petals). (a) Estuary, rising tide, (b) Estuary, ebb tide, (c) Saguenay River, rising tide, (d) Saguenay River, ebb tide. Petal length indicates relative frequency.

Counter current movements were rare. The areas showing the highest level of consistency were corridor 3 along the southern shore of Île aux Lièvres, corridor 4 and 7 off the Batture aux Alouettes, and corridor 8 off Île aux Basques.

Mean water depths ( $\pm$  SD) were  $59 \pm 63$  m ( $n = 450$  cells) and  $92 \pm 102$  m ( $n = 927$  cells) for corridors and non-corridors, respectively (Fig. 8). The kernel density curves and mean ( $\pm$  95% CI) depth distributions differed significantly between corridors and non-corridors (Fig. 8; t-test on log-transformed data:  $t = 4.11$ ,  $df = 1053$ ,  $P < 0.001$ ). Distributions for corridors and non-corridors were similar for the most part, with a mode near 25 m and a long right tail. However, non-corridors showed a second mode near 320 m. The greatest depth value associated with a point sampling session was 328 m.

## DISCUSSION

Our study presented an innovative approach to habitat connectivity and the quantitative assessment of habitat use that is applicable to any terrestrial or aquatic species. Our framework where we analyzed direct observational data with two independent approaches confirmed the existence of an elaborate network of transit corridors in the core of the summer range of SLE beluga. Combining these connectivity results with survey data allowed to estimate the minimum proportion of the population likely to transit along the corridors (i.e., the corridor connective value), to use each of the delineated concentration areas, and to be exposed to local stressors.

The corridor network within the SLE beluga summer range consisted in a central confluence

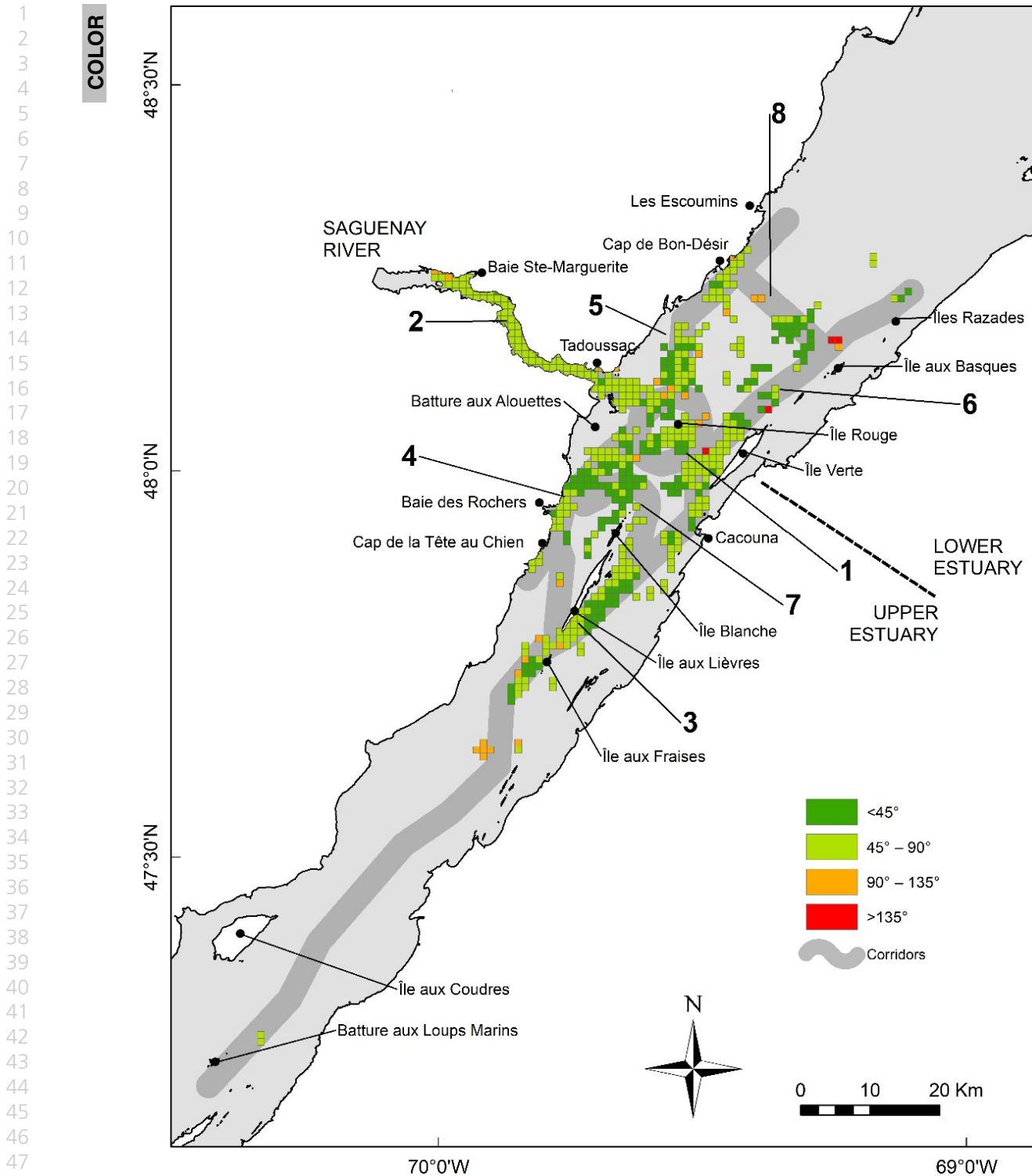


Fig. 7. Absolute difference (from  $0^\circ$  to  $180^\circ$ ) between step direction and simultaneous surface current direction averaged over every strong and contiguous vectors. In cells with multiple strong vectors, the results from the vector with the smallest average difference were shown. Also shown are the corridors labeled from 1 to 8.

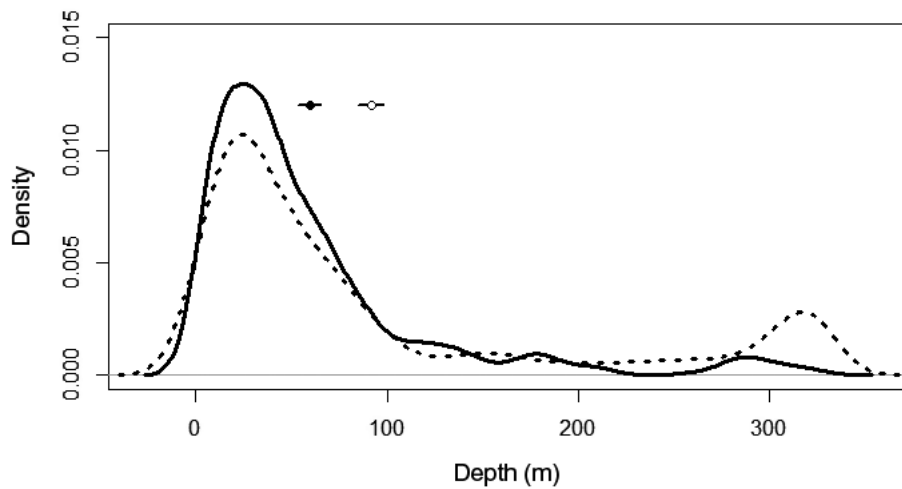


Fig. 8. Kernel density curves and mean ( $\pm$  95% CI) of depth in cells belonging to a corridor (solid line, filled symbol) and cells not belonging to a corridor (dashed line, open symbol). Whether cells belong to corridors or not is in accordance with the highest level of severity. The Saguenay River and depths  $<2$  m were not included.

zone connected with three main branches to the Upper and Lower Estuary, and to the Saguenay River. Although an increased coverage of the summer range extremes is needed, the network appears to reach almost every end of the summer range, which is consensually recognized as Bature aux Loups Marins, Rimouski, and Baie Sainte-Marguerite (Mosnier et al. 2010, Gosselin et al. 2017). The majority of herd movements were performed along the main axis of either the Estuary or the Saguenay River, leading to most corridors being oriented roughly parallel to the nearest coast. Notable exceptions were corridors 1, 7, 8 (Fig. 5) which provided some connectivity among concentration areas across the Estuary. These results confirm the existence of a high level of connectivity among the SLE beluga concentration areas. The existence of transit corridors in the SLE beluga summer range was proposed by Lefebvre et al. (2012) based on travel speed profiles of beluga herds, although the authors did not speculate on their location.

Using a multi-level severity approach to corridor delineation provided the flexibility but also the robustness to acknowledge the possibility that a corridor route crossed a concentration area or data-poor areas. In the first case, vector amplitude would be typically low, and the steps associated with transiting herds would be masked by the lack of directionality in movement by

resident herds. In the second case, a lack of strong vector in that specific area would be explained by a low sampling effort.

The spatial segregation pattern we found is consistent with previous observations for this population (Michaud 1993) and with reports from multiple investigators where females with calves used shallower waters and ventured further upstream in estuaries (Smith et al. 1994, Loseto et al. 2006). Hypotheses for sex- and age-based spatial segregation in odontocetes in general and in beluga in particular were reviewed by Michaud (2005) and Loseto et al. (2006). They include avoidance of predation and intraspecific competition, age-specific or sex-specific forage selection and energy budget, and social preference. While sex cannot be determined accurately from field observations, biopsy samples collected over the years from SLE beluga confirmed that groups of white adults are predominantly composed of males only, and that white adults in groups where juveniles and calves are present are females (Michaud and Lesage, *personal observation*). Based on this assumption, herds of females with calves and/or juveniles would use the shallower, warmer, and brackish Upper Estuary, typical of a nearshore environment or an estuary proper. In the Lower Estuary, these segments of the population would also use waters near the south shore and on the southern slope

1 of the LC, but rarely the deeper and colder mar-  
2 ine waters of the Laurentian Channel which were  
3 used mostly by herds of adult males only. Corri-  
4 dor 8, which ran across the Lower Estuary, was  
5 particularly interesting given the heterogeneity  
6 in herd compositions described for this general  
7 sector. All types of herds likely transited via cor-  
8 ridor 8 (Figs. 2, 5), although movements of white  
9 adults across the Estuary axis were clearly more  
10 extensive than those of herds with juveniles or  
11 calves as they reached the deeper waters of the  
12 LC.

13 The configuration of the corridor network sug-  
14 gests that a large proportion of the population is  
15 susceptible to visit the Île Rouge area. Also, all  
16 population segments are likely to visit this latter  
17 area along with the Saguenay River, and, to a les-  
18 ser extent, the south shore of the Lower Estuary.  
19 In a conservation perspective, this also indicates  
20 that a large portion of the population could be  
21 exposed to any habitat degradation or stressor  
22 located in those areas, even stationary ones with  
23 a spatially restricted reach. A consequence of  
24 sex- and age-based spatial segregation is that all  
25 segments of the population are not equally vul-  
26 nerable to a stressor, assuming it is localized with  
27 a spatial extent of effects less than the full sum-  
28 mer range of the population. A stressor located  
29 in the Upper Estuary would mostly impact  
30 females, calves, and juveniles, whereas a stressor  
31 in the Laurentian Channel would impact mainly  
32 adult males.

33 The corridor network overlaps the St. Law-  
34 rence Seaway, a busy seaway with an average  
35 5000-6000 transits annually (McQuinn et al.  
36 2011), and which crosses the entire beluga sum-  
37 mer range, mostly along the north shore of the  
38 Estuary, thereby exposing all age and sex classes.  
39 A secondary seaway branches off at the Sague-  
40 nay River mouth and reaches the upper Sague-  
41 nay River. Also, the Saguenay River mouth is  
42 considered the noisiest and busiest concentration  
43 area for SLE beluga (McQuinn et al. 2011, Ger-  
44 vaise et al. 2012), as a consequence of whale-  
45 watching activities, recreational boating, and the  
46 simultaneous operation of two or three ferries  
47 depending on season and time of day. The loca-  
48 tion of the corridors, and degree of connectivity  
49 among concentration areas, indicate that a large  
50 proportion of the individuals in the population  
51 might be exposed to noise pollution and risks of

collision. Collision risk with merchant ships is  
not considered problematic for beluga, but colli-  
sions with small crafts have been reported in the  
past (Lair et al. 2016). Both merchant ships and  
small craft can interfere with the normal behav-  
ior of beluga and reduce their acoustic space,  
with potential consequences on foraging effi-  
ciency, health, and social bonds (Lesage et al.  
1999, Gervaise et al. 2012, Williams et al. 2017).  
Previous studies examining beluga exposure to  
noise in several of their habitat, and the potential  
benefits of mitigation measures reducing noise  
exposure emphasized the existence of noise  
refuge in the beluga summer range (McQuinn  
et al. 2011, Lesage et al. 2014, Chion et al. 2017,  
DFO 2018). These areas of lesser exposure to  
noise were located in the Saguenay River, and in  
the southern portion of the Estuary where  
islands in the center of the Estuary acted as a bar-  
rier to the noise issued from traffic along the  
north shore. While we cannot exclude that some  
spatial structure in site fidelity exists within  
regions used by a specific type of herd, the NW-  
SE connectivity documented both in the Upper  
and Lower Estuary suggests that no segment of  
the population is likely to be fully protected from  
noise exposure by constraining their domain to  
these noise refuges.

Herds transiting in corridors showed a strong  
propensity to move upstream during rising tide  
and downstream during ebb tide in the Estuary.  
The pattern was not as evident in the Saguenay  
River. Also, herd movement directions were  
highly consistent with local surface current direc-  
tions. This is consistent with the energy land-  
scape principle which should lead animals to  
move at the best possible cost:benefit trade-off  
(Shepard et al. 2013). This includes reducing  
costs of locomotion between habitat patches by  
taking advantage of the current, which consti-  
tutes in the marine environment the equivalent  
of the well-documented avian behavior of  
migrating with tailwinds. Situations where bel-  
uga synchronized their movements with current  
schedule were reported in Alaska and Canadian  
Arctic populations (Smith et al. 1994, EzEr et al.  
2008). In both studies, the beluga took advantage  
of flooding tidal currents when transiting to shal-  
low waters at the mouth of tributaries. The study  
area exhibits a highly complex hydrology. The St.  
Lawrence Estuary belongs to the salt-wedge

1 type, and the lower Saguenay River included in  
 2 our study area is a fiord (El-Sabh and Silverberg  
 3 1990). This means that both waterways are  
 4 strongly stratified marine environments. Conse-  
 5 quently, the tidal current during rising tide, com-  
 6 posed of marine water, flows upstream  
 7 underneath the fresh water layer. This typical  
 8 estuarine circulation results in weaker current at  
 9 the surface during rising tide than during ebb  
 10 tide (Saucier and Chassé 2000). This tidal phe-  
 11 nomenon may explain the contrast we found in  
 12 distribution of current data between tidal phases.

13 The cells belonging to a corridor were located  
 14 on average in shallower waters than cells not  
 15 belonging to a corridor. This result may indicate  
 16 that a selection process is occurring on the part  
 17 of transiting beluga herds. The difference in  
 18 water depth is mainly due to the larger represen-  
 19 tation of the non-corridors in the deepest waters  
 20 (>200 m) of the Laurentian Channel. But we  
 21 question the meaning of this result because,  
 22 despite its great depths exerting a strong influ-  
 23 ence on the bathymetric profile, the actual rela-  
 24 tive importance of the Laurentian Channel  
 25 (depths > 200 m) in the study area was small  
 26 with only 13% of relative surface and 8% of the  
 27 sampling effort. This leaves perhaps little possi-  
 28 bility to detect the existence of a corridor in that  
 29 area and allows us to conclude that, for the most  
 30 part, there is no difference in water depth  
 31 between cells belonging to a corridor and cells  
 32 not belonging to a corridor. Excluding the  
 33 Laurentian Channel from the dataset, the  
 34 difference in water depth between corridors and  
 35 non-corridors vanishes completely ( $49 \pm 4$  m,  
 36  $n = 429$  and  $48 \pm 3$  m,  $n = 762$  for corridors and  
 37 non-corridors, respectively).

38 Our study indicates that beluga use currents  
 39 when transiting but depth does not seem to be a  
 40 driver in shaping habitat selection process and  
 41 routes of the transit corridors. The amount of  
 42 connectivity we found is probably an underesti-  
 43 mation of total connectivity since the sampling  
 44 effort was low near the ends of the study area,  
 45 and therefore, focal follows were limited. Addi-  
 46 tional observations through photo-identification  
 47 work or radio or satellite tracking of individual  
 48 whales might reveal more transit corridors con-  
 49 necting the CAs in those areas, as well as a sub-  
 50 structure in site fidelity and spatial segregation  
 51 within the summer distribution range.

## ACKNOWLEDGMENTS

The authors wish to thank field observers from GREMM (Renaud Pintiaux, Tim Ferrero) and from DFO (Yves Morin, Sébastien Lemieux Lefebvre, Caroline Guimont). Denis Lefavre and Alain D'Astous provided data on surface current and Jean-François Gosselin granted access to aerial survey data. We are also grateful to Benoît Bruneau, Cédric Juillet, Arnaud Mosnier, Martin-Hugues Saint-Laurent, and Yvan Simard for advices and thoughtful discussions. Funding was provided by the Species at Risk program from DFO. We also thank one anonymous reviewer for suggested improvements to the manuscript. VL conceived the ideas; VL, MM, RM collected the data; JFO analyzed the data; JFO and VL wrote the manuscript.

## LITERATURE CITED

- Bryan-Brown, D. N., C. J. Brown, J. M. Hughes, and R. M. Connolly. 2017. Patterns and trends in marine population connectivity research. *Marine Ecology Progress Series* 585:243–256.
- Caliński, T., and J. Harabasz. 1974. A dendrite method for cluster analysis. *Communications in Statistics-Theory and Methods* 3:1–27.
- Canadian Hydrographic Service. 2018. Tides, Current, and Water Levels database. <http://www.tides.gc.ca/eng/find/zone/2>
- Caron, L. M. J., and T. G. Smith. 1990. Philopatry and site tenacity of belugas, *Delphinapterus leucas*, hunted by the Inuit at the Nastapoka estuary, eastern Hudson Bay. *Canadian Bulletin of Fisheries and Aquatic Sciences* 224:69–79.
- Chadenet, V. 1997. Fréquentation et bilan d'activité du béluga, *Delphinapterus leucas*, du Saint-Laurent dans la Baie Sainte-Marguerite. Thesis. Université Laval, Québec, Québec, Canada.
- Chion, C., D. Lagrois, J. Dupras, S. Turgeon, I. H. McQuinn, R. Michaud, N. Ménard, and L. Parrott. 2017. Underwater acoustic impacts of shipping management measures: results from a social-ecological model of boat and whale movements in the St. Lawrence River Estuary (Canada). *Ecological Modelling* 354:72–87.
- Colbeck, G. J., P. Duchesne, L. D. Postma, V. Lesage, M. O. Hammill, and J. Turgeon. 2013. Groups of related belugas (*Delphinapterus leucas*) travel together during their seasonal migrations in and around Hudson Bay. *Proceedings of the Royal Society B: Biological Sciences* 280:20122552.
- COSEWIC [Committee on the Status of Endangered Wildlife in Canada]. 2014. COSEWIC assessment and status report on the Beluga Whale

- 1 *Delphinapterus leucas*, St. Lawrence Estuary,  
2 Canada. COSEWIC, Ottawa, Ontario, Canada.
- 3 Cushman, S. A., B. McRae, F. Adriaensen, P. Beier, M.  
4 Shirley, and K. Zeller. 2013. Biological corridors  
5 and connectivity. Pages 384–404 in D. W. Macdonald  
6 and K. J. Willis, editors. Key topics in conserva-  
7 tion biology 2. Wiley-Blackwell, Hoboken, New  
8 Jersey, USA.
- 9 DFO [Department of Fisheries and Oceans]. 2018.  
10 Potential effects of the construction of marine ter-  
11 minals in the Saguenay fjord on the St. Lawrence  
12 beluga whale and its habitat. DFO Canadian  
13 Science Advisory Secretariat, Science Response  
14 2018/025.
- 15 El-Sabh, M. I., and N. Silverberg. 1990. Oceanography  
16 of a large-scale estuarine system: the St. Lawrence.  
17 Voluem 39. Springer-Verlag, New York, New York,  
18 USA.
- 19 EzEr, T., R. Hobbs, and L.-Y. Oey. 2008. On the move-  
20 ment of beluga whales in Cook Inlet, Alaska: simu-  
21 lations of tidal and environmental impacts using a  
22 hydrodynamic inundation model. *Oceanography*  
23 21:186–195.
- 24 Fall, A., M.-J. Fortin, M. Manseau, and D. O'Brien.  
25 2007. Spatial graphs: principles and applications  
26 for habitat connectivity. *Ecosystems* 10:448–461.
- 27 Gervaise, C., Y. Simard, N. Roy, B. Kinda, and N.  
28 Menard. 2012. Shipping noise in whale habitat:  
29 characteristics, sources, budget, and impact on belu-  
30 gas in Saguenay–St. Lawrence Marine Park hub. *Journal of the Acoustical Society of America*  
31 132:76–89.
- 32 Gosselin, J.-F., M. O. Hammill, and A. Mosnier. 2014.  
33 Summer abundance indices of St. Lawrence Estu-  
34 ary beluga (*Delphinapterus leucas*) from a photo-  
35 graphic survey in 2009 and 28 line transect surveys  
36 from 2001 to 2009. DFO Canadian Science Advi-  
37 sory Secretariat Research Document 2014/021.
- 38 Gosselin, J.-F., M. O. Hammill, A. Mosnier, and V.  
39 Lesage. 2017. Abundance index of St. Lawrence  
40 Estuary beluga, *Delphinapterus leucas*, from aerial  
41 visual surveys flown in August 2014 and an  
42 update on reported deaths. DFO Canadian Science  
43 Advisory Secretariat Research Document 2017/019.
- 44 Henein, K., and G. Merriam. 1990. The elements of  
45 connectivity where corridor quality is variable.  
46 *Landscape Ecology* 4:157–170.
- 47 Hobbs, R. J. 1992. The role of corridors in conservation:  
48 solution or bandwagon? *Trends in Ecology and*  
49 *Evolution* 7:389–392.
- 50 Kaufman, L., and P. J. Rousseeuw. 1990. Finding  
51 groups in data: An introduction to cluster analysis.  
John Wiley & Sons, New York, New York, USA.
- Kindlmann, P., and F. Burel. 2008. Connectivity mea-  
sures: a review. *Landscape Ecology* 23:879–890.
- Lair, S., L. Measures, and D. Martineau. 2016. Patho-  
logic findings and trends in mortality in the beluga  
(*Delphinapterus leucas*) population of the St Lawr-  
ence Estuary, Quebec, Canada, from 1983 to 2012.  
*Veterinary Pathology* 53:22–36.
- Lefebvre, S. L., V. Lesage, R. Michaud, and M. M.  
Humphries. 2018. Classifying and combining  
herd surface activities and individual dive pro-  
files to identify summer behaviours of beluga  
(*Delphinapterus leucas*) from the St. Lawrence  
Estuary. Canada. *Canadian Journal of Zoology*  
96:393–410.
- Lefebvre, S. L., R. Michaud, V. Lesage, and D. Ber-  
teaux. 2012. Identifying high residency areas of the  
threatened St. Lawrence beluga whale from fine-  
scale movements of individuals and coarse-scale  
movements of herds. *Marine Ecology Progress Ser-  
ies* 450:243–257.
- Lesage, V., C. Barrette, M. C. Kingsley, and B. Sjare.  
1999. The effect of vessel noise on the vocal behav-  
ior of belugas in the St. Lawrence River estuary,  
Canada. *Marine Mammal Science* 15:65–84.
- Lesage, V., L. N. Measures, A. Mosnier, S. Lair, R.  
Michaud, and P. Béland. 2014. Mortality patterns in  
St. Lawrence Estuary beluga (*Delphinapterus leu-  
cas*), inferred from the carcass recovery data, 1983–  
2012. *Canadian Science Advisory Secretariat,*  
*Research Document,* 2013/118.
- Loseto, L., P. Richard, G. Stern, J. Orr, and S. Ferguson.  
2006. Segregation of Beaufort Sea beluga whales  
during the open-water season. *Canadian Journal of*  
*Zoology* 84:1743–1751.
- Mardia, K. V., and P. E. Jupp. 2009. *Directional statis-  
tics.* John Wiley & Sons, New York, New York,  
USA.
- McQuinn, I. H., V. Lesage, D. Carrier, G. Larrivée, Y.  
Samson, S. Chartrand, R. Michaud, and J. Theri-  
ault. 2011. A threatened beluga (*Delphinapterus leu-  
cas*) population in the traffic lane: vessel-generated  
noise characteristics of the Saguenay–St. Lawrence  
Marine Park, Canada. *Journal of the Acoustical*  
*Society of America* 130:3661–3673.
- McRae, B. H., B. G. Dickson, T. H. Keitt, and V. B.  
Shah. 2008. Using circuit theory to model connec-  
tivity in ecology, evolution, and conservation. *Ecol-  
ogy* 89:2712–2724.
- Michaud, R. 1993. Distribution estivale du béluga du  
Saint-Laurent; synthèse. 1986–1992. *Rapport tech-  
nique canadien des sciences halieutiques et aqua-  
tiques* 1906.
- Michaud, R. 2005. Sociality and ecology of the odonto-  
cetes. Pages 303–326 in K. E. Ruckstuhl and P. Neu-  
haus, editors. *Sexual segregation in vertebrates:  
ecology of the two sexes.* Cambridge University  
Press, Cambridge, UK.

- 1 Monteith, K. L., M. M. Hayes, M. J. Kauffman, H. E.  
2 Copeland, and H. Sawyer. 2018. Functional attrib-  
3 utes of ungulate migration: Landscape features  
4 facilitate movement and access to forage. *Ecologi-  
5 cal Applications* 28:2153–2164.
- 6 Mosnier, A., T. Doniol-Valcroze, J.-F. Gosselin, V.  
7 Lesage, L. N. Measures, and M. O. Hammill. 2015.  
8 Insights into processes of population decline using  
9 an integrated population model: the case of the St.  
10 Lawrence Estuary beluga (*Delphinapterus leucas*).  
11 *Ecological Modelling* 314:15–31.
- 12 Mosnier, A., R. Larocque, M. Lebeuf, J.-F. Gosselin, S.  
13 Dubé, V. Lapointe, V. Lesage, D. Lefavre, S. Sen-  
14 neville, and C. Chion. 2016. Définition et caractéri-  
15 sation de l'habitat du béluga (*Delphinapterus leucas*)  
16 de l'estuaire du Saint-Laurent selon une approche  
17 écosystémique. DFO Canadian Science Advisory  
18 Secretariat, Research Document 2016/052.
- 19 Mosnier, A., V. Lesage, J.-F. Gosselin, S. Lemieux  
20 Lefebvre, M. O. Hammill, and T. Doniol-Valcroze.  
21 2010. Information relevant to the documentation of  
22 habitat use by St. Lawrence beluga (*Delphinapterus  
23 leucas*), and quantification of habitat quality. DFO  
24 Canadian Science Advisory Secretariat, Research  
25 Document 2009/098.
- 26 O'Corry-Crowe, G., R. Suydam, L. Quakenbush, B.  
27 Potgieter, L. Harwood, D. Litovka, T. Ferrer, J.  
28 Citta, V. Burkanov, and K. Frost. 2018. Migratory  
29 culture, population structure and stock identity in  
30 North Pacific beluga whales (*Delphinapterus leucas*).  
31 *PLOS ONE* 13:e0194201.
- 32 O'Corry-Crowe, G., R. Suydam, L. Quakenbush, T. G.  
33 Smith, C. Lydersen, K. M. Kovacs, J. Orr, L. Har-  
34 wood, D. Litovka, and T. Ferrer. 2020. Group struc-  
35 ture and kinship in beluga whale societies.  
36 *Scientific Reports* 10:1–21.
- 37 Pendoley, K. L., G. Schofield, P. A. Whittock, D. Ierodia-  
38 conou, and G. C. Hays. 2014. Protected species use of  
39 a coastal marine migratory corridor connecting marine  
40 protected areas. *Marine Biology* 161:1455–1466.
- 41 R Development Core Team. 2019. R: a language and  
42 environment for statistical computing. R Founda-  
43 tion for Statistical Computing, Vienna, Austria.
- 44 Roeber, C. L., R. J. Van Aarde, and K. Leggett. 2013.  
45 Functional connectivity within conservation  
46 networks: delineating corridors for African ele-  
47 phants. *Biological Conservation* 157:128–135.
- 48 Saucier, F. J., and J. Chassé. 2000. Tidal circulation and  
49 buoyancy effects in the St. Lawrence Estuary,  
50 Canada. *Atmosphere-Ocean* 38:1–52.
- 51 Saucier, F. J., J. Chassé, M. Couture, A. D'Astous, R.  
Dorais, D. Lefavre, and A. Gosselin. 1999. The  
making of a surface current atlas of the St. Lawr-  
ence. Pages 87–97 in C. A. Brebbia and P. Anagnos-  
topoulos, editors. *Transactions on the Built  
Environment*. Volume 43. WIT Press, Southamp-  
ton, UK.
- Sawyer, H., C. W. LeBeau, T. L. McDonald, W. Xu, and  
A. D. Middleton. 2019. All routes are not created  
equal: An ungulate's choice of migration route can  
influence its survival. *Journal of Applied Ecology*  
56:1860–1869.
- Shepard, E. L., R. P. Wilson, W. G. Rees, E. Grundy, S.  
A. Lambertucci, and S. B. Vosper. 2013. Energy  
landscapes shape animal movement ecology.  
*American Naturalist* 182:298–312.
- Smith, T., M. Hammill, and A. Martin. 1994. Herd  
composition and behaviour of white whales (*Del-  
phinapterus leucas*) in two Canadian arctic estuar-  
ies. *Meddelelser om Grønland. BioScience* 39:175–  
184.
- Tibshirani, R., G. Walther, and T. Hastie. 2001. Estimat-  
ing the number of clusters in a data set via the  
gap statistic. *Journal of the Royal Statistical  
Society: Series B (Statistical Methodology)* 63:  
411–423.
- Turgeon, J., P. Duchesne, G. J. Colbeck, L. D. Postma,  
and M. O. Hammill. 2012. Spatiotemporal segrega-  
tion among summer stocks of beluga (*Delphi-  
napterus leucas*) despite nuclear gene flow:  
implication for the endangered belugas in eastern  
Hudson Bay (Canada). *Conservation Genetics*  
13:419–433.
- Williams, R., R. C. Lacy, E. Ashe, A. Hall, C. Lehoux,  
V. Lesage, I. McQuinn, and S. Plourde. 2017. Pre-  
dicting responses of St. Lawrence beluga to envi-  
ronmental change and anthropogenic threats to  
orient effective management actions. DFO Cana-  
dian Science Advisory Secretariat Research Docu-  
ment 2017/027.

## SUPPORTING INFORMATION

Additional Supporting Information may be found online at: <http://onlinelibrary.wiley.com/doi/10.1002/ecs2.3560/full>